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**Real-Time Determination of Swine Manure Liquid Compost Maturity by
Oxidation-Reduction Potential and pH Monitoring**

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Abstract

The aim of this study was to develop a real-time diagnostic method for assessing the maturity of liquid compost from swine manure by monitoring the oxidation-reduction potential (ORP) and pH, two key indicators of the oxidative status of organic matter and nitrogen compounds. The experimental system comprised a cylindrical bioreactor equipped with an aeration and mixing device. During the liquid composting process of swine manure in the bioreactor, the ORP and pH were continuously monitored, while changes in the liquid compost maturity and concentrations of soluble nitrogen and organic matter were observed. In addition, the physicochemical differences between mature liquid compost (MLC) and non-mature liquid compost (NMLC) were comparatively analyzed. The MLC exhibited significantly higher ORP and lower pH with lower ammonium nitrogen concentrations and higher concentrations of oxidized nitrogen compounds ($\text{NO}_x\text{-N}$) compared to NMLC. A diagnostic method combining ORP and pH thresholds was proposed, where samples with an ORP greater than or equal to 180 mV were classified as MLC. In addition, samples with an ORP between 0 and 180 mV and $\text{pH} < 8$ were classified as MLC. Field validation at a swine manure liquid composting facility confirmed the accuracy of the diagnostic method, achieving 100% agreement with liquid compost maturity analysis using an automatic liquid compost maturity tester. This study provides a simple and effective tool for optimizing liquid compost production, minimizing ammonia emissions, and enhancing liquid compost maturity in swine manure recycling systems.

Keywords: swine manure, liquid compost, maturity, real-time determination, oxidation-reduction potential, pH

Introduction

Swine manure has long been used as an organic fertilizer to supply nitrogen to agricultural systems, and the recycling of nutrients in livestock manure plays a significant role in enhancing agricultural sustainability [1–4]. Because of its rich content of essential nutrients and minerals required for crop growth, swine manure has been reported to influence soil fertility positively [5–12]. However, the expansion of the livestock industry has led to an increase in the amount of manure generated, resulting in various environmental concerns. Livestock manure, which contains high concentrations of nitrogen, phosphorus, and organic matter, can cause severe contamination of soil and water systems if not properly managed [13–15]. Additionally, ammonia (NH_3) emitted during the land application of swine manure can degrade air quality and contribute to environmental pollution through atmospheric

deposition, causing soil acidification and eutrophication [16–18]. NH_3 is of particular concern because it can be converted into ammonium aerosols in the atmosphere, forming fine particulate matter that poses a potential threat to public health and results in significant social health costs [19, 20]. In response, regulatory frameworks have been adopted in many countries [21].

Another concern when applying swine manure to agricultural land is the potential phytotoxicity of ammonium nitrogen (NH_4^+) to crops. NH_4^+ toxicity in plants has been extensively studied and is known to adversely affect plant growth [22, 23]. Classical and modern hypotheses regarding the mechanisms of NH_4^+ toxicity were summarized by Esteban et al. [24]. Although studies on the relationship between soil nitrogen forms and plant growth have shown that sensitivity to ammonium-N ($\text{NH}_4\text{-N}$) varies among plant species, nitrate-N ($\text{NO}_3\text{-N}$) is generally considered the preferred nitrogen form in most higher plants [25, 26].

Aerobic treatment is considered the most effective method for reducing odors and preserving nitrogen in a more beneficial form when swine manure is used as an organic fertilizer. Most odorous compounds produced during swine manure decomposition, such as nitrogen, sulfur, volatile fatty acids, and aromatic compounds, are by-products of anaerobic microbial activity [27]. The odor-reducing effect of aerobic treatment is well-established, and its basic principle is to provide sufficient dissolved oxygen to aerobic bacteria to actively degrade odorous substances [28]. Additionally, aerobic treatment can convert ammonium nitrogen ($\text{NH}_4\text{-N}$) in swine manure into nitrate nitrogen ($\text{NO}_3\text{-N}$), a form with low phytotoxicity, and is considered the most efficient method for such transformation [29]. Owing to these advantages, several countries, including South Korea and Japan, have long preferred to apply swine manure to soil only after aerobic treatment in the form of liquid compost [30, 31]. In particular, the Ministry of Environment of South Korea introduced the concept of maturity of liquid compost derived from aerobically treated swine manure to ensure environmental safety and mandated maturity testing prior to land application.

However, owing to the high variability of swine manure, it is difficult to determine the exact duration of aerobic treatment required to achieve sufficient maturity. Moreover, the real-time assessment and diagnosis of liquid compost maturity are currently unavailable, making it challenging for farms to apply swine manure in an environmentally sustainable form. A method that can accurately and easily determine the appropriate timing for land application by enabling real-time monitoring of liquid compost maturity is needed to address this issue. Therefore, this study was conducted to develop a method for the real-time diagnosis of liquid compost maturity during aerobic treatment using oxidation-reduction potential (ORP) and pH, which are widely used indicators of oxidation levels in aqueous environments.

Materials and Methods

Experimental Setup

As shown in Fig. 1, a liquid composting system was constructed to investigate the changes in maturity during the swine manure composting process and to analyze the behavior of the oxidation-reduction potential (ORP) and pH across different maturity levels. The bioreactor was fabricated using a cylindrical acrylic vessel with a total volume of 30 L, a diameter of 400 mm, and a height of 350 mm. The system was equipped with a mixer (MS-17GY; JEIO TECH, Daejeon, Korea), air stones, a flow meter (RMA-14-SSV; Dwyer Instruments, Michigan City, USA), and a blower (BT-6500; Chuang Xing Manufacturing Co. Ltd., Tsuen Wan, China) to provide aerobic conditions. Two air stones were installed symmetrically at the bottom of the bioreactor to supply air, and the airflow was independently regulated using individual flow meters.

Probes to monitor the changes in ORP and pH during the liquid composting process were installed at the mid-water level of the bioreactor. The signals were transmitted to a programmable logic controller (PLC) and were subsequently recorded on a computer. The ORP and pH probes were validated by measuring standard solutions at the start of each experimental cycle. When deviations were detected, immediate recalibration was conducted. Additionally, regular accuracy checks were performed on a weekly basis to ensure the reliability of sensor measurements throughout the monitoring period.

Raw swine manure was introduced into the bioreactor via an influent pump, and the amounts of inflow and outflow were controlled by adjusting the pump operation time. An effluent port was installed in the middle of the effective water height in the reactor and connected to an effluent pump, which was used to discharge the liquid compost and collect the samples. The experiments were conducted in a controlled laboratory environment maintained at 25 °C to minimize the influence of temperature fluctuations on the liquid composting process.

Operational Condition and Procedure

An experiment was conducted to reflect the typical operational procedures used in liquid composting facilities for swine manure. In such facilities, it is a common practice to retain a sufficient number of microorganisms (activated sludge) within the bioreactor while discharging only a portion of the mature liquid compost, followed by the introduction of fresh raw manure. This approach ensured efficient liquid composting. This study adopted the same

operational practice: prior to the inflow of swine manure, a volume of liquid compost equivalent to the target influent volume was discharged, and raw swine manure was introduced into the reactor while retaining the remaining liquid compost. This approach allowed activated microorganisms to be retained within the bioreactor, thereby supporting stable microbial activity throughout the experiment.

The operating conditions of the bioreactors used in this study are listed in Table 1. The reactor had an effective volume of 25 L and was operated under continuous aeration at an airflow rate of 0.05 L air/L reactor volume/min following the introduction of raw manure. In Korea, the aeration rate of 0.03 m³ air/m³ reactor volume/min is generally applied for liquid composting, following "Standard Design Guide for Livestock Manure Resource Recovery Facilities". However, due to the shallow depth (350 mm) of the experimental bioreactor used in this study, reduced oxygen transfer efficiency was anticipated. To compensate for this limitation, the aeration rate was increased by approximately 70%, resulting in a final applied rate of 0.05 m³ air/m³ reactor volume/min. Each liquid composting cycle was continued until all NH₄-N introduced via the influent was oxidized. The cycle was terminated once NH₄-N was fully depleted, and the next experimental cycle was initiated (any overlapping data collected after the complete oxidation of NH₄-N were excluded from the analysis). The volume of raw swine manure added to each cycle was adjusted to achieve loading rates ranging from 5% (1.25 L) to 40% (10 L). As mentioned above, each experimental cycle was maintained until complete oxidation of NH₄-N was achieved and lasted approximately 4 to 19 days, depending on the level of hydraulic loading and reaction efficiency. A total of 67 samples were collected at 1–2 day intervals during the composting process to evaluate changes in maturity and composition.

[Insert Table 1 here]

The initial liquid compost and raw manure used in the experiment were collected from a swine farm in Hoengseong-gun, Gangwon-do, South Korea. Because aeration could not be performed during transportation and preparation, the collected initial liquid compost was aerated in the laboratory for one week prior to use to preserve microbial activity. Raw swine manure was collected after solid separation using a solid-liquid separator (MCL-300D; MUHAN TECHNICAL Co. Ltd., Daegu, Korea) and then frozen until use. The characteristics of the liquid compost and raw manure are listed in Table 2.

[Insert Table 2 here]

Monitoring System

ORP (SOR 400G; Samsan Korea, Daegu Metropolitan City, Korea) and pH sensors (SPH-200G; Samsan Korea) were installed at the mid-level height in the bioreactor working volume to monitor changes in the oxidation-reduction potential (ORP) and pH during the liquid composting process. Voltage signals generated from each sensor were transmitted through the corresponding ORP (CRN-96 ORP; Samsan Korea) and pH instruments (CRN-96 pH; Samsan Korea), converted to current signals of 4–24 mA, and relayed to a programmable logic controller (PLC, XGB, LS ELECTRIC Co., Gyeonggi-do, Korea). Sensor readings were measured at one-second intervals, and the average of every 60 data points (i.e., one-minute intervals) was recorded and stored on a computer connected to the PLC.

Analysis of Liquid Compost Composition and Maturity

This study primarily focused on the analysis of soluble nitrogen and organic matter based on the principle that aerobic conditions promote the oxidation of these components during liquid composting. Additional analyses of total Kjeldahl nitrogen (TKN) and solids were conducted according to standard methods. The growth of activated sludge in the bioreactor significantly affected the solid fraction, limiting the relevance of solid-based indicators in evaluating liquid compost maturity. Therefore, soluble nitrogen and organic carbon concentrations were primarily used to interpret the maturity of the liquid compost.

The collected samples were immediately centrifuged at 3,000 rpm for 10 min and then filtered using Whatman No. 6 filter paper. The filtrate was analyzed for soluble total organic carbon (STOC) and soluble nitrogen components. STOC was measured using a total organic carbon analyzer (Torch; Teledyne Tekmar, Mason, OH, USA), and $\text{NH}_4\text{-N}$ and $\text{NO}_x\text{-N}$ concentrations were analyzed using an automated water quality analyzer (QuickChem 8600; Lachat, Weston, CT, USA). The concentration of free NH_3 in each sample was estimated based on the equilibrium ratio of NH_4^+ and NH_3 at a given pH and temperature following the method proposed by Emerson et al. [32]. The pH value measured at the time of sampling was used, and a temperature of 20 °C—the temperature at which maturity was measured—was applied.

The maturity of the liquid compost samples was evaluated using an automated liquid compost maturity analyzer (LMQ2000; KOREA SPECTRAL PRODUCTS, Seoul, Korea). This device assesses compost maturity based on gas emissions (NH_3 and H_2S) and sample absorbance, with 97.4% agreement with seed germination test results [33].

Statistical Analysis

Some indicators for mature and non-mature liquid compost do not satisfy the assumption of normality. Therefore, the non-parametric Mann–Whitney U test was employed to compare the distributions of ORP, pH, and concentrations of soluble nitrogen and organic matter between the two groups. Statistical analyses were performed using the SAS software. Statistical significance was set at $p < 0.05$.

Results and Discussion

Physicochemical Changes and Dynamic Characteristic of Liquid Composting Process

To investigate the physicochemical changes within the bioreactor during the liquid composting process, the variations in ORP, pH, and soluble nitrogen species across each experimental cycle—operated under different hydraulic loading rates—are presented in Fig. 2. Additionally, the time required to complete nitrification under each loading condition is summarized in Table 3 to analyze the effect of loading variation on $\text{NH}_4\text{-N}$ oxidation efficiency. (Results from all cycles are not shown due to the repetitive nature of the trends observed.) As shown in Fig. 2 A–D, despite the differences in hydraulic loading, all cycles exhibited a common pattern characterized by a decrease in $\text{NH}_4\text{-N}$ concentration and a corresponding increase in $\text{NO}_x\text{-N}$, accompanied by a rise in ORP and a decline in pH as aeration progressed. The consistent pattern observed supports the utility of ORP and pH as effective real-time indicators for evaluating the maturity of liquid compost.

On the other hand, the time required to complete nitrification did not show a direct proportional relationship with the initial hydraulic loading conditions. For example, although cycles B and C were both operated at a 30% hydraulic loading rate, the time required for complete nitrification differed significantly, at 3.9 and 18.8 days, respectively (Fig. 2). Notably, even cycle D, which was operated at a higher 40% hydraulic loading rate, reached nitrification completion in just 5.4 days, suggesting that the result from cycle C and D deviates from the intuitive expectation that increased influent volume would result in longer nitrification periods. This discrepancy indicates that actual organic and nitrogen loading, rather than hydraulic loading alone, must be considered when interpreting the results.

Table 3 summarizes the $\text{NH}_4\text{-N}$ and STOC loading rates along with the corresponding nitrification durations for each cycle. In cycles A, B, and D, as the hydraulic loading increased from 20% to 40%, the $\text{NH}_4\text{-N}$ and STOC loadings also increased proportionally. In contrast, cycle C, despite having a 30% hydraulic loading rate, exhibited the highest $\text{NH}_4\text{-N}$ and STOC loadings at 913.4 mg/L and 3749.7 mg/L, respectively, implying that the raw swine

manure concentration was substantially higher than in the other cycles. As a result, cycle C required the most oxygen for the oxidation of organic matter and nitrogen, leading to the longest nitrification period among all tested conditions.

[Insert Table 3 here]

These findings strongly support the necessity of developing real-time diagnostic technologies for assessing compost maturity. Given the considerable variability in nitrogen and organic content in swine manure and the impracticality of visual or sensory evaluation in field conditions, a reliable and on-site method for maturity assessment is essential for effective management of liquid composting systems.

Characteristics of Mature Liquid Compost (MLC) and Non-matured Liquid Compost (NMLC)

Among the 67 samples analyzed, 27 were classified as MLC, and 40 as NMLC. The differences in characteristics between the two groups are summarized in Table 4. Significant differences were observed in the concentrations of soluble nitrogen and organic carbon between the MLC and NMLC samples, with the MLC exhibiting a notably higher level of oxidation. In particular, the difference in the composition of soluble nitrogen was pronounced: the concentrations of $\text{NH}_4\text{-N}$ and $\text{NO}_x\text{-N}$ in MLC averaged 55.4 ± 57.8 mg/L and 450.3 ± 195.9 mg/L, respectively, indicating a predominance of oxidized nitrogen forms. In contrast, NMLC samples exhibited much higher concentrations of reduced nitrogen, with $\text{NH}_4\text{-N}$ and $\text{NO}_x\text{-N}$ levels of 516.5 ± 222.6 mg/L and 116.6 ± 177.7 mg/L, respectively. Soluble organic carbon was also lower in MLC (2260.9 ± 746.6 mg/L) compared to NMLC (3206.0 ± 1079.7 mg/L). As a result, the MLC samples, which were characterized by more advanced oxidation of nitrogen and organic matter, exhibited higher ORP values and lower pH values than NMLC, with statistically significant differences. The ORP in MLC averaged 133.3 ± 91.4 mV, compared to -65.1 ± 127.5 mV in NMLC. Similarly, the pH of MLC was lower (7.6 ± 0.4) than that of NMLC (8.3 ± 0.3). Consequently, the potential for ammonia volatilization from MLC was markedly lower than that from NMLC. The calculated free NH_3 concentrations, estimated based on measured $\text{NH}_4\text{-N}$ levels, pH, and temperature, were 1.3 ± 1.7 mg/L for MLC and 47.7 ± 29.8 mg/L for NMLC, indicating a substantially reduced risk of NH_3 emissions from MLC. These results suggested that managing liquid compost maturity not only effectively mitigates NH_3 volatilization but also preserves nitrogen in the more plant-beneficial $\text{NO}_x\text{-N}$ form.

[Insert Table 4 here]

Overall, compared with NMLC, MLC was characterized by lower concentrations of $\text{NH}_4\text{-N}$ and STOC, higher $\text{NO}_x\text{-N}$ levels, elevated ORP, and reduced pH. This pattern aligns with findings from previous studies comparing MLC and NMLC compositions. Jeon et al. analyzed 150 liquid compost samples collected from the Liquid Compost Supply Center and classified them by maturity [34, 35]. The MLCs showed lower levels of organic matter and $\text{NH}_4\text{-N}$, higher $\text{NO}_x\text{-N}$ concentrations, higher ORP, and lower pH than the NMLCs, which agrees with the results of the present study. However, the ORP values measured in this study— 133.3 ± 91.4 mV for MLC and -65.1 ± 127.5 mV for NMLC—were notably higher than those reported by Jeon et al. This discrepancy may be due to methodological differences; the present study measured ORP in real-time within the bioreactor, whereas Jeon et al. measured ORP and pH in laboratory settings after sample collection from the Liquid Compost Supply Center.

The distribution patterns of $\text{NH}_4\text{-N}$, $\text{NO}_x\text{-N}$, STOC, ORP, and pH for both the MLC and NMLC are presented in Fig. 3. The most distinct differences between the two groups were found in nitrogen-related parameters. Clear distinctions in the distribution ranges and median values of the $\text{NH}_4\text{-N}$ and $\text{NO}_x\text{-N}$ concentrations were evident. Because MLCs exhibited both lower $\text{NH}_4\text{-N}$ levels and lower pH, the resulting free NH_3 concentrations were significantly lower than those in NMLC. Although STOC was also lower in MLC and statistically different from NMLC, its distribution range overlapped between the groups, making it a less reliable differentiator than the nitrogen components. Importantly, the higher oxidation states of nitrogen and organic matter in the MLC were reflected in the distribution patterns of ORP and pH. These findings supported the feasibility of developing a rapid and intuitive diagnostic tool for evaluating the maturity status of liquid compost based on real-time ORP and pH monitoring.

Maturity Characteristics of Liquid Compost According to ORP Distribution

All 67 samples were categorized into four ORP (mV) ranges (i.e., $\text{ORP} \geq 150$ mV, $150 > \text{ORP} \geq 0$ mV, $0 > \text{ORP} \geq -150$ mV, and $\text{ORP} < -150$ mV) to evaluate the feasibility of predicting liquid compost maturity in real-time based on ORP measurements during the liquid composting process. The pH, $\text{NH}_4\text{-N}$, $\text{NO}_x\text{-N}$, STOC, and free NH_3 values were analyzed for each range, as summarized in Table 5.

[Insert Table 5 here]

All samples within the $\text{ORP} \geq 150$ mV range were classified as MLC when the samples were grouped by ORP range. In this group, the average concentrations of $\text{NH}_4\text{-N}$ and $\text{NO}_x\text{-N}$ were 13.1 mg/L and 420.5 mg/L, respectively, indicating a high level of nitrogen oxidation. Consequently, the pH was 7.36 on average, and the estimated free NH_3

concentration calculated based on $\text{NH}_4\text{-N}$ levels and pH was as low as 0.1 mg/L, suggesting almost no potential for odor emissions. In contrast, all samples were determined to be NMLC in the range where $\text{ORP} < 0$ mV. The proportion of $\text{NH}_4\text{-N}$ in the soluble nitrogen fraction was relatively high in this group, with pH values > 8 . The estimated free NH_3 concentration was between 53.5 and 54.5 mg/L, indicating a high potential for NH_3 volatilization. Meanwhile, only 53.6% of the samples were classified as MLC in the intermediate range of $150 > \text{ORP} \geq 0$ mV, meaning that matured and non-matured samples were present in similar proportions, making it difficult to clearly distinguish maturity status based solely on ORP. This result is in agreement with early research by Charpentier et al. who reported that ORP levels vary depending on the behavior and internal reactions of dissolved oxygen, organic matter, nitrogen, and phosphorus in aqueous systems [36]. According to their findings, when ORP falls below -300 mV, reduction of organic matter and nitrogen dominates; near 0 mV, organic matter is oxidized but nitrification is limited; and above $+100$ mV, both organic matter and nitrogen are actively oxidized. These findings support the results of our study, which showed that in the ORP range between 0 and 150 mV, nitrification remains limited, residual $\text{NH}_4\text{-N}$ concentrations are high, and the distinction between matured and immature samples becomes unclear.

To better understand the characteristics of liquid compost in the ambiguous range of $150 > \text{ORP} \geq 0$ mV, the physicochemical properties of both MLC and NMLC within this range were compared (Table 6). Even within this range, the MLC showed a higher proportion of $\text{NO}_x\text{-N}$ in the soluble nitrogen fraction, whereas the NMLC showed a higher proportion of $\text{NH}_4\text{-N}$. However, average ORP values were very similar: 62.1 ± 40.7 mV for MLC and 66.0 ± 56.5 mV for NMLC. Additionally, the ORP ranges (minimum to maximum) did not differ significantly between the two groups, indicating that ORP alone could not be used to differentiate maturity in this interval. Interestingly, the pH values differed significantly. Within the $150 > \text{ORP} \geq 0$ mV range, the average pH of MLC was 7.77 ± 0.18 (ranging from 7.37 to 8.13), while NMLC showed a higher average pH of 8.32 ± 0.18 (ranging from 7.94 to 8.63).

[Insert Table 6 here]

Overall, these results suggested that monitoring the ORP during the liquid composting process allows for the prediction of maturity. Furthermore, in cases where ORP alone does not clearly distinguish between MLC and NMLC, pH monitoring can serve as an effective complementary indicator for real-time maturity diagnosis.

Maturity Characteristics of Liquid Compost According to pH Distribution

Similar to the ORP, the feasibility of using real-time pH monitoring to assess the maturity of liquid compost was evaluated by classifying the samples based on pH ranges, as shown in Table 7. It is well established that biological nitrification consumes alkalinity, thereby lowering pH. In this study, a decrease in pH was associated with a reduced proportion of $\text{NH}_4\text{-N}$ and a corresponding increase in $\text{NO}_x\text{-N}$ in the total soluble nitrogen content. As a result, all samples with $\text{pH} \leq 7.5$ exhibited very low $\text{NH}_4\text{-N}$ concentrations, with $\text{NO}_x\text{-N}$ accounting for the majority of the soluble nitrogen, and were all classified as MLC. Conversely, in samples with $8.0 < \text{pH} \leq 8.5$, the rate of MLC dropped to just 8%, and all samples with $\text{pH} > 8.5$ were identified as NMLC. The concentration of free NH_3 , calculated based on the pH and $\text{NH}_4\text{-N}$ concentration of each sample, averaged only 0.1 mg/L in samples with $\text{pH} \leq 7.5$. However, this value increased sharply to 41.34 mg/L in the $8.0 < \text{pH} \leq 8.5$ range and further to 75.1 mg/L in samples with $\text{pH} > 8.5$. These findings underscore the importance of producing MLC, not only for reducing atmospheric NH_3 emissions from agricultural soils, but also for managing odor-related complaints.

[Insert Table 7 here]

The MLC rate was 72.7% in the intermediate range of $7.50 < \text{pH} \leq 8.00$, indicating reduced discriminatory power for accurate diagnosis. Therefore, the physicochemical characteristics of the MLC and NMLC samples within this pH range were compared (Table 8). Within this range, the composition of soluble nitrogen ($\text{NH}_4\text{-N}$ and $\text{NO}_x\text{-N}$) differed markedly between MLC and NMLC, resulting in a corresponding difference in free NH_3 concentrations. This indicates a clear compositional distinction based on maturity. The mean pH values of MLC and NMLC in this range were nearly identical, at 7.79 ± 0.12 and 7.80 ± 0.15 , respectively, indicating that pH alone could not reliably predict compost maturity. In contrast, the ORP values differed significantly between MLC and NMLC, averaging 85.1 ± 64.1 mV and -162.4 ± 110.6 mV, respectively. This suggested that in pH ranges in which compost maturity is difficult to determine, incorporating ORP monitoring can enhance diagnostic reliability.

[Insert Table 8 here]

Real-Time Diagnosis of Liquid Compost Maturity Based on ORP and pH Monitoring

Physicochemical changes occurring during the liquid composting process resulted in notable differences in nitrogen and organic matter concentrations between MLC and NMLC. Consequently, both the mean values and distribution ranges of the ORP and pH varied between the two groups. ORP and pH demonstrated clear thresholds that could distinguish MLC from NMLC, indicating their potential as effective real-time indicators of compost maturity. However, when used independently, each indicator had limitations in clearly differentiating between MLC

and NMLC within certain ranges. Integrating both indicators may improve the reliability of real-time maturity diagnostics.

The distributions of ORP and pH for MLC and NMLC were analyzed to validate this hypothesis and are presented in Fig. 4. The distribution ranges of ORP and pH differed between the MLC and NMLC. Most MLC samples were distributed within the range of $\text{ORP} \geq 0 \text{ mV}$ and $\text{pH} \leq 8$, with only one outlier exhibiting ORP 246.3 mV and pH 8.01. It is widely known that higher oxidation levels in water typically result in elevated ORP and decreased pH. However, previous studies have shown that pH may increase after complete oxidation of $\text{NH}_4\text{-N}$ due to CO_2 stripping [37, 38]. Based on this, the single MLC sample falling outside the proposed diagnostic range of $\text{ORP} \geq 0 \text{ mV}$ and $\text{pH} \leq 8$ likely underwent complete nitrification prior to sampling, with the elevated pH presumably resulting from the stripping of CO_2 . This sample had an $\text{NH}_4\text{-N}$ concentration of 0 mg/L and a $\text{NO}_x\text{-N}$ concentration of 182.1 mg/L. This suggested that when the ORP is sufficiently high, pH may not be a critical factor in maturity diagnostics. Notably, no NMLC samples were observed with $\text{ORP} \geq 150 \text{ mV}$.

While most samples within the range of $\text{ORP} \geq 0 \text{ mV}$ and $\text{pH} \leq 8$ were classified as MLC, one sample was identified as NMLC. This sample had ORP 19.5 mV and pH 7.9, with $\text{NH}_4\text{-N}$ and $\text{NO}_x\text{-N}$ concentrations of 217.9 and 436.2 mg/L, respectively. Although $\text{NO}_x\text{-N}$ was higher, the residual $\text{NH}_4\text{-N}$ concentration was considerably greater than the MLC average (Table 4), and the pH was relatively high. Consequently, the free NH_3 concentration reached 7.9 mg/L, which likely led to an immature classification. These findings imply that when diagnosing maturity based on the criteria of $\text{ORP} \geq 0 \text{ mV}$ and $\text{pH} \leq 8$ if both indicators are near their threshold values, the organic matter and nitrogen in the liquid compost may not have been fully oxidized. Therefore, samples within this range may still be immature, and caution should be exercised.

Four diagnostic criteria were developed based on these results using a combination of ORP and pH, as summarized in Table 9. The first criterion, $\text{ORP} \geq 180 \text{ mV}$, indicated MCL (dotted line in Fig. 4), and pH was not considered. This was based on the finding that a sufficiently high ORP alone may reliably indicate maturity. In this study, all samples with $\text{ORP} \geq 150 \text{ mV}$ were classified as MLC. A safety margin of 20% was added to 150 mV to ensure a more conservative threshold, resulting in 180 mV being used as the final criterion. All samples in this range were classified as MLC, with $\text{NH}_4\text{-N}$ accounting for only 2.7% of the total soluble nitrogen, and the free NH_3 concentration averaging only 0.14 mg/L, indicating a minimal risk of NH_3 volatilization.

[Insert Table 9 here]

The second criterion, $180 > \text{ORP} \geq 0$ and $\text{pH} \leq 8$, represented near-matured liquid compost. Although there was a slight risk of misclassification, the agreement rate was 95%. Within this range, $\text{NH}_4\text{-N}$ constituted 16.6% of the total soluble nitrogen, and the average free NH_3 concentration was 2.19 mg/L. While oxidation levels were slightly lower than in the $\text{ORP} \geq 180$ mV group, they were still significantly better than in NMLC. However, when both the ORP and pH values were close to the threshold, there was a higher chance of misclassification in this range. Operators may need to wait for the ORP to increase further or for the pH to decrease before confirming the maturity, and some level of professional judgment may be required.

If the ranges of $\text{ORP} \geq 180$ mV and $180 > \text{ORP} \geq 0$ with $\text{pH} \leq 8$ can be considered indicators of MLC, all other ranges should be regarded as NMLC, as they are likely to contain high levels of $\text{NH}_4\text{-N}$ and exhibit elevated pH values, increasing the risk of NH_3 volatilization. Although these liquid composts are immature, real-time diagnostic results can inform operational adjustments such as increasing aeration or reducing the influent load to promote liquid compost maturation more effectively.

Validation of the Liquid Compost Maturity Diagnostic Method

Field testing was conducted at a swine manure liquid composting facility in Samseong-myeon, Eumseong-gun, Korea, to evaluate the reliability of the real-time liquid compost maturity diagnostic method that combines ORP and pH. During the liquid composting process, ORP and pH were continuously monitored, and the maturity of the liquid compost was assessed. The results are shown in Fig. 5. Of the 31 samples analyzed to verify diagnostic reliability, 22 were classified as MLC and 9 as NMLC. All samples identified as matured satisfied the diagnostic criteria proposed in this study, which were defined as either $\text{ORP} \geq 180$ mV or $180 > \text{ORP} \geq 0$ mV and $\text{pH} \leq 8$. None of the NMLC cases was mistakenly classified as MLC. Interestingly, the pH of MLC collected from the field liquid composting facility was notably lower than those observed under laboratory conditions, averaging 6.6 ± 0.5 , with the lowest value reaching 6.0. This difference may be attributed to the field composting system operating under higher organic and nitrogen loading rates, resulting in more advanced nitrification or the accumulation of $\text{NO}_x\text{-N}$ over prolonged operational periods.

In conclusion, the diagnostic method developed through laboratory experiments demonstrated high accuracy for the real-time evaluation of liquid compost maturity under field conditions. The diagnostic accuracy was 100% for all 31 field samples tested. Therefore, the combined monitoring of ORP and pH presents a practical and effective approach for managing the liquid composting process and improving compost quality at pig farms. Further

accumulation of field data and additional studies are expected to enhance the precision and applicability of this diagnostic technique.

Conclusion

The aim of this study was to develop a real-time diagnostic method for assessing the maturity of swine manure liquid compost using oxidation-reduction potential (ORP) and pH, two key indicators that reflect the oxidative status of organic matter and nitrogen during the liquid composting process. Through comprehensive monitoring and analysis of 67 compost samples, significant differences in nitrogen and organic matter concentrations were observed between the MLC and NMLC. These differences were consistently reflected in the ORP and pH values, thereby establishing their potential as reliable indicators of liquid compost maturity.

Liquid composts with $\text{ORP} \geq 180 \text{ mV}$ were consistently classified as matured, while those with $180 > \text{ORP} \geq 0 \text{ mV}$ under $\text{pH} \leq 8$ showed a maturity classification accuracy of 95%, thereby validating the effectiveness of the proposed combined diagnostic criteria. In contrast, compost outside these thresholds exhibited elevated $\text{NH}_4\text{-N}$ concentrations and free NH_3 levels, indicating an increased risk of ammonia volatilization and environmental odor issues. The practical applicability of the diagnostic method was further validated on a commercial-scale swine farm, where 100% diagnostic accuracy was achieved across 31 field samples.

These findings suggested that the combined use of ORP and pH enables the intuitive and accurate real-time assessment of liquid compost maturity, thereby supporting better operational control and product quality in liquid composting systems. Further data accumulation and extended field validation are expected to enhance the precision of this diagnostic approach and provide a valuable tool for sustainable nutrient recycling in agriculture.

References

1. Burton GW. Tifton 85 Bermuda grass—early history of its creation, selection, and evaluation. *Crop Sci.* 2001;41:5–6. <https://doi.org/10.2135/cropsci2001.4115>
2. Freitas JAS, Silva VR, Luz FB, Kaiser DR, Zwirtes AL. Soil carbon and physical-mechanical properties after successive applications of swine and poultry organic waste. *Pesqui Agropecu Trop.* 2018;48:390–8. <https://doi.org/10.1590/1983-40632018v4852412>

3. Guardini R, Comin JJ, Schmitt DE, Tiecher T, Bender MA, Dos Santos DR, et al. Accumulation of phosphorus fractions in typic Hapludalf soil after long-term application of pig slurry and deep pig litter in a no-tillage system. *Nutr Cycl Agroecosyst*. 2012;93:215–25. <https://doi.org/10.1007/s10705-012-9511-3>
4. da Silva Oliveira DM, de Lima RP, Barreto MSC, Verburg EEJ, Mayrink GCV. Soil organic matter and nutrient accumulation in areas under intensive management and swine manure application. *J Soils Sediments*. 2017;17:1–10. <https://doi.org/10.1007/s11368-016-1474-6>
5. Boitt G, Schmitt DE, Gatiboni LC, Wakelin SA, Black A, Sacomori W, et al. Fate of phosphorus applied to soil in pig slurry under cropping in southern Brazil. *Geoderma*. 2018;321:164–72. <https://doi.org/10.1016/j.geoderma.2018.02.010>
6. Formentini TA, Mallmann FJK, Pinheiro A, Fernandes CVS, Bender MA, Da Veiga M, et al. Copper and zinc accumulation and fractionation in a clayey Hapludox soil subject to long-term pig slurry application. *Sci Total Environ*. 2015;536:831–9. <https://doi.org/10.1016/j.scitotenv.2015.07.110>
7. Grohskopf MA, Correa JC, Cassol PC, Nicoloso RS, Fernandes DM. Copper and zinc forms in soil fertilized with pig slurry in the bean crop. *Rev Bras Eng Agric Ambient*. 2016;20:823–9. <https://doi.org/10.1590/1807-1929/agriambi.v20n9p823-829>
8. Hountin JA, Karam A, Couillard D, Cescas MP. Use of a fractionation procedure to assess the potential for P movement in a soil profile after 14 years of liquid pig manure fertilization. *Agric Ecosyst Environ*. 2000;78:77–84. [https://doi.org/10.1016/S0167-8809\(99\)00112-7](https://doi.org/10.1016/S0167-8809(99)00112-7)
9. Johannes A, Matter A, Schulin R, Weiskopf P, Baveye PC, Boivin P. Optimal organic carbon values for soil structure quality of arable soils. Does clay content matter? *Geoderma*. 2017;302:14–21. <https://doi.org/10.1016/j.geoderma.2017.04.021>
10. Lourenzi CR, Ceretta CA, Tiecher TL, Lorensini F, Cancian A, Stefanello L, et al. Forms of phosphorus transfer in runoff under no-tillage in a soil treated with successive swine effluents applications. *Environ Monit Assess*. 2015;187:209. <https://doi.org/10.1007/s10661-015-4437-2>
11. Scheid DL, Silva RF, Silva VR, Ros COD, Pinto MAB, Gabriel M, et al. Changes in soil chemical and physical properties in pasture fertilised with liquid swine manure. *Sci Agric*. 2020;77:e20190017. <https://doi.org/10.1590/1678-992X-2019-0017>
12. Won S, You BG, Shim S, Ahmed N, Choi YS, Ra C. Nutrient variations from swine manure to agricultural land. *Asian-Australas J Anim Sci*. 2018;31:763–72. <https://doi.org/10.5713/ajas.17.0634>

13. Folino A, Zema DA, Calabrò PS. Organic matter removal and ammonia recovery by optimised treatments of swine wastewater. *J Environ Manage.* 2020;270:110692. <https://doi.org/10.1016/j.jenvman.2020.110692>
14. Panetta DM, Powers WJ, Lorimor JC. Management strategy impacts on ammonia volatilization from swine manure. *J Environ Qual.* 2005;34:1119–30. <https://doi.org/10.2134/jeq2004.0313>
15. Viancelli A, Kunz A, Steinmetz RLR, Kich JD, Souza CK, Canal CW, et al. Performance of two swine manure treatment systems on chemical composition and on the reduction of pathogens. *Chemosphere.* 2013;90:1539–44. <https://doi.org/10.1016/j.chemosphere.2012.08.055>
16. Behera SN, Sharma M, Aneja VP, Balasubramanian R. Ammonia in the atmosphere: a review on emission sources, atmospheric chemistry and deposition on terrestrial bodies. *Environ Sci Pollut Res Int.* 2013;20:8092–131. <https://doi.org/10.1007/s11356-013-2051-9>
17. Ellis RA, Murphy JG, Markovic MZ, VandenBoer TC, Makar PA, Brook J, et al. The influence of gas-particle partitioning and surface-atmosphere exchange on ammonia during BAQS-met. *Atmos Chem Phys.* 2011;11:133–45. <https://doi.org/10.5194/acp-11-133-2011>
18. Ti C, Gao B, Luo Y, Wang X, Wang S, Yan X. Isotopic characterization of $\text{NH}_x\text{-N}$ in deposition and major emission sources. *Biogeochemistry.* 2018;138:85–102. <https://doi.org/10.1007/s10533-018-0432-3>
19. Ge X, Schaap M, Kranenburg R, Segers A, Reinds GJ, Kros H, et al. Modeling atmospheric ammonia using agricultural emissions with improved spatial variability and temporal dynamics. *Atmos Chem Phys Discuss.* 2020; 2020:1–51. <https://doi.org/10.5194/acp-20-16055-2020>
20. Ti C, Xia L, Chang SX, Yan X. Potential for mitigating global agricultural ammonia emission: a meta-analysis. *Environ Pollut.* 2019;245:141–8. <https://doi.org/10.1016/j.envpol.2018.10.124>
21. Khoshnevisan B, Duan N, Tsapekos P, Awasthi MK, Liu Z, Mohammadi A, et al. A critical review on livestock manure biorefinery technologies: sustainability, challenges, and future perspectives. *Renew Sustain Energy Rev.* 2021;135:110033. <https://doi.org/10.1016/j.rser.2020.110033>
22. Coskun D, Britto DT, Li M, Becker A, Kronzucker HJ. Rapid ammonia gas transport accounts for futile transmembrane cycling under $\text{NH}_3/\text{NH}_4^+$ toxicity in plant roots. *Plant Physiol.* 2013;163:1859–67. <https://doi.org/10.1104/pp.113.225961>
23. Bittsánszky A, Pilinszky K, Gyulai G, Komives T. Overcoming ammonium toxicity. *Plant Sci.* 2015;231:184–90. <https://doi.org/10.1016/j.plantsci.2014.12.005>

24. Esteban R, Ariz I, Cruz C, Moran JF. Review: Mechanisms of ammonium toxicity and the quest for tolerance. *Plant Sci.* 2016;248:92–101. <https://doi.org/10.1016/j.plantsci.2016.04.008>
25. Britto DT, Kronzucker HJ. NH_4^+ toxicity in higher plants: a critical review. *J Plant Physiol.* 2002;159:567–84. <https://doi.org/10.1078/0176-1617-0774>
26. Wang ZH, Miao YF, Li SX. Effect of ammonium and nitrate nitrogen fertilizers on wheat yield in relation to accumulated nitrate at different depths of soil in drylands of China. *Field Crops Res.* 2015;183:211–24. <https://doi.org/10.1016/j.fcr.2015.07.019>
27. Rappert S, Müller R. Odor compounds in waste gas emissions from agricultural operations and food industries. *Waste Manag.* 2005;25:887–907. <https://doi.org/10.1016/j.wasman.2005.07.008>
28. Zhu J. A review of microbiology in swine manure odor control. *Agric Ecosyst Environ.* 2000;78:93–106. [https://doi.org/10.1016/S0167-8809\(99\)00116-4](https://doi.org/10.1016/S0167-8809(99)00116-4)
29. Riaño B, García-González MC. On-farm treatment of swine manure based on solid-liquid separation and biological nitrification–denitrification of the liquid fraction. *J Environ Manage.* 2014;132:87–93. <https://doi.org/10.1016/j.jenvman.2013.10.014>
30. Haga K. Animal waste problems and their solution from the technological point of view in Japan. *Jpn Agric Res.* 1998;32:203–10
31. Won S, Ahmed N, You BG, Shim S, Kim SS, Ra C. Nutrient production from Korean poultry and loading estimations for cropland. *J Anim Sci Technol.* 2018;60:3. <https://doi.org/10.1186/s40781-018-0160-1>
32. Emerson K, Russo RC, Lund RE, Thurston RV. Aqueous ammonia equilibrium calculations: effect of pH and temperature. *J Fish Res Bd Can.* 1975;32:2379–83. <https://doi.org/10.1139/f75-274>
33. Park J. A Study on the Establishment of Liquid Fertilizer Germination Test and Mechanical Maturity Assessment Methods [Master's thesis]. Gangwon-do, KR: Kangwon National University; 2018. https://dcollection.kangwon.ac.kr/public_resource/pdf/000000029961_20250615200316.pdf
34. Jeon SJ, Kim SR, Rho KS, Choi DY, Kim DK, Lee MG. Physicochemical characteristics of liquid fertilizer made from pig manure in Korea. *J Anim Environ Sci.* 2012;18:221–8
35. Jeon SJ, Kim SR, Hong IG, Kim HJ, Kim DG, Lee MG. A comparative study on correlation through physiochemical property comparison of livestock liquid fertilizer. *J Anim Environ Sci.* 2013;19:163–8. <https://doi.org/10.11109/JAES.2013.19.2.163>

36. Charpentier J, Florentz M, David G. Oxidation-reduction potential (ORP) regulation: a way to optimize pollution removal and energy savings in the low load activated sludge process. *Wat Sci Tech.* 1987;19:645-55. <https://doi.org/10.2166/wst.1987.0244>
37. Won SG, Ra CS. Biological nitrogen removal with a real-time control strategy using moving slope changes of pH(mV)- and ORP-time profiles. *Water Res.* 2011;45:171–8. <https://doi.org/10.1016/j.watres.2010.08.030>
38. Kim S, Reza A, Shim S, Won S, Ra C. Development of a real-time controlled bio-liquor circulation system for swine farms: A lab-scale study. *Animals (Basel).* 2021;11:311. <https://doi.org/10.3390/ani11020311>

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Tables and Figures

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Table 1. Operational conditions for the bioreactor

Item	Value and condition
Working volume (L)	25
Hydraulic loading rate based on reactor working volume (%)	5–40
Operation mode	Batch type
Aeration mode	Continuously aeration
Aeration rate (L/L/min)	0.05

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Table 2. Characteristics of initial liquid compost and swine manure

Item	Unit	Liquid compost	Swine manure
TS		10400.0	17950.0
TVS		4577.8	11016.7
TSS		5455.6	8211.1
TVSS		4044.4	7022.2
NH ₄ -N	mg/L	Not detected	2681.9
NO _x -N		182.1	Not detected
TKN		732.8	3933.8
T-N		915.2	3933.8
STOC		1991.7	11222.8

TS, total solids; TVS, total volatile solids; TSS, total suspended solids; TVSS, total volatile suspended solids;

TKN, total Kjeldahl nitrogen; T-N, total nitrogen; STOC, soluble total organic carbon

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Table 3. Comparison of nitrification completion time under varying hydraulic, organic, and nitrogen loading rate

Experimental cycle	Loading rate (mg/L)			Elapsed time for complete nitrification (d)
	Hydraulic (%)	NH ₄ -N (mg/L)	STOC (mg/L)	
A	20	406.2	1094.9	2.8
B	30	629.9	1718.9	3.9
C	30	914.3	3749.7	18.8
D	40	863.5	2200.8	5.4

Table 4. Comparison of characteristics between matured and non-matured liquid compost

Item	Matured (<i>n</i> = 27)	Non-matured (<i>n</i> = 40)	<i>p</i> -value
NH ₄ -N (mg/L)	55.4 ± 57.8	516.5 ± 222.6	< 0.001
NO _x -N (mg/L)	450.3 ± 195.9	116.6 ± 177.7	< 0.001
STOC (mg/L)	2260.9 ± 746.6	3206.0 ± 1079.7	0.006
ORP (mV)	133.3 ± 91.4	-65.1 ± 127.5	< 0.001
pH	7.6 ± 0.4	8.3 ± 0.3	< 0.001
Free NH ₃ (mg/L)	1.3 ± 1.7	47.7 ± 29.8	< 0.001

516 Table 5. Characteristics of liquid composts according to ORP level

Range of ORP (mV)		pH	NH ₄ -N (mg/L)	NO _x -N (mg/L)	STOC (mg/L)	Free NH ₃ (mg/L)	No. of samples	Percent of matured (%)
ORP ≥150	Avg.	7.36	13.1	420.5	1906.5	0.1	12	100
	Std.	0.44	9.1	165.6	391.3	0.2		
	Min.	6.77	0.0	161.5	1388.2	0.0		
	Max.	8.01	28.2	580.4	2543.5	0.8		
150> ORP ≥0	Avg.	8.03	218.5	335.1	2909.4	17.4	28	53.6
	Std.	0.34	175.9	246.1	1042.8	23.5		
	Min.	7.37	7.2	0.04	1420.8	0.2		
	Max.	8.63	577.6	839.4	4569.8	81.0		
0> ORP ≥-150	Avg.	8.42	524.8	141.1	2991.4	54.5	16	0
	Std.	0.20	124.2	217.4	904.4	25.3		
	Min.	7.84	346.1	0.0	1925.5	11.7		
	Max.	8.71	791.7	724.3	4536.8	113.8		
-150> ORP	Avg.	8.20	688.7	12.4	3476.2	53.5	11	0
	Std.	0.38	280.7	40.8	1248.3	37.1		
	Min.	7.56	198.5	0.0	1812.0	5.7		
	Max.	8.67	1168.4	135.4	5274.4	105.6		

526 Table 6. Characteristics of matured and non-matured liquid compost in the range of $150 > \text{ORP (mV)} \geq 0$

Maturity of samples		ORP (mV)	pH	NH ₄ -N (mg/L)	NO _x -N (mg/L)	STOC (mg/L)	Free NH ₃ (mg/L)	No. of samples
Matured	Avg.	62.1	7.77	95.5	474.2	2583.0	2.8	15
	Std.	40.7	0.18	74.0	219.9	858.7	3.6	
	Min.	11.4	7.37	7.2	46.9	1420.8	0.2	
	Max.	131.4	8.13	286.5	839.4	3959.0	14.6	
Non-matured	Avg.	66.0	8.32	360.4	174.5	3208.5	34.3	13
	Std.	56.5	0.21	149.9	166.6	1140.3	25.5	
	Min.	1.5	7.94	183.3	0.0	1526.2	7.3	
	Max.	140.4	8.63	577.6	589.1	4569.8	81.0	

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528 Table 7. Characteristics of liquid compost according to pH level

Range of pH		ORP (mV)	NH ₄ -N (mg/L)	NO _x -N (mg/L)	STOC (mg/L)	Free NH ₃ (mg/L)	No. of samples	matured percent (%)
pH ≤ 7	Avg.	217.8	17.1	572.7	2292.1	0.05	3	100
	Std.	55.0	6.6	6.7	0.3	0.01		
	Min.	164.4	11.6	568.6	2291.9	0.04		
	Max.	274.3	24.5	580.4	2292.3	0.06		
7.00 < pH ≤ 7.50	Avg.	218.7	15.6	473.6	1899.3	0.11	6	100
	Std.	66.8	10.1	92.5	639.1	0.09		
	Min.	124.5	0.8	393.2	1388.2	0.01		
	Max.	292.4	30.5	601.6	2977.2	0.28		
7.50 < pH ≤ 8.00	Avg.	17.6	174.9	377.7	2496.2	4.06	22	72.7
	Std.	136.3	215.4	273.6	793.0	4.27		
	Min.	-267.7	1.3	0.0	1420.8	0.05		
	Max.	209.8	833.3	839.4	3959.0	11.99		

8.00 < pH ≤ 8.50	Avg.	-45.9	476.6	142.8	3153.3	41.34	25	8
	Std.	136.2	270.3	160.1	1142.6	25.92		
	Min.	-346.8	0.00	0.0	1526.2	0.00		
	Max.	246.3	1168.4	589.1	5274.4	95.10		
pH < 8.50	Avg.	-19.1	576.4	34.9	3369.1	75.14	11	0
	Std.	119.7	90.9	52.0	1110.2	28.68		
	Min.	-230.3	461.9	0.0	1938.0	53.66		
	Max.	137.6	743.4	154.7	4569.8	113.84		

Table 8. Characteristics of matured and non-matured liquid compost in the range of $7.50 < \text{pH} \leq 8.00$

Maturity of samples		ORP (mV)	pH	NH ₄ -N (mg/L)	NO _x -N (mg/L)	STOC (mg/L)	Free NH ₃ (mg/L)	No. of samples
Matured	Avg.	85.1	7.79	72.3	446.7	2444.9	1.80	16
	Std.	64.1	0.12	56.3	231.2	870.9	1.70	
	Min.	11.4	7.55	1.3	46.9	1420.8	0.05	
	Max.	209.8	7.97	213.5	839.4	3959.0	6.38	
Non-matured	Avg.	-162.4	7.80	448.4	193.5	2619.1	10.08	6
	Std.	110.6	0.15	248.9	313.2	634.6	2.84	
	Min.	-267.7	7.56	198.5	0.0	1812.0	5.67	
	Max.	19.5	7.94	833.3	724.3	3351.5	11.99	

538 Table 9. Characteristics of liquid compost according to diagnostic criteria of oxidation-reduction potential (ORP)
 539 and pH

Parameter	Diagnostic criteria of liquid compost maturity			
	ORP ≥ 180	180 > ORP ≥ 0 and pH ≤ 8	180 > ORP ≥ 0 and pH > 8	ORP < 0
Matured rate (%)	100	95	7.7	0
Number of samples	9	18	13	27
NH ₄ -N (mg/L)	11.8 \pm 9.8	78.6 \pm 63.5	365.7 \pm 145.6	581.9 \pm 225.2
NO _x -N (mg/L)	424.2 \pm 164.4	472.7 \pm 207.8	161.7 \pm 150.4	99.9 \pm 191.1
Total soluble nitrogen (NH ₄ -N+NO _x -N, mg/L)	436.0	551.3	527.4	681.8
Total soluble organic carbon (mg/L)	2296.1 \pm 1177.9	2531.7 \pm 847.1	3101.5 \pm 1185.0	3258.0 \pm 1062.4
Free NH ₃ (mg/L)	0.14 \pm 0.27	2.19 \pm 2.30	37.12 \pm 24.59	57.62 \pm 29.59

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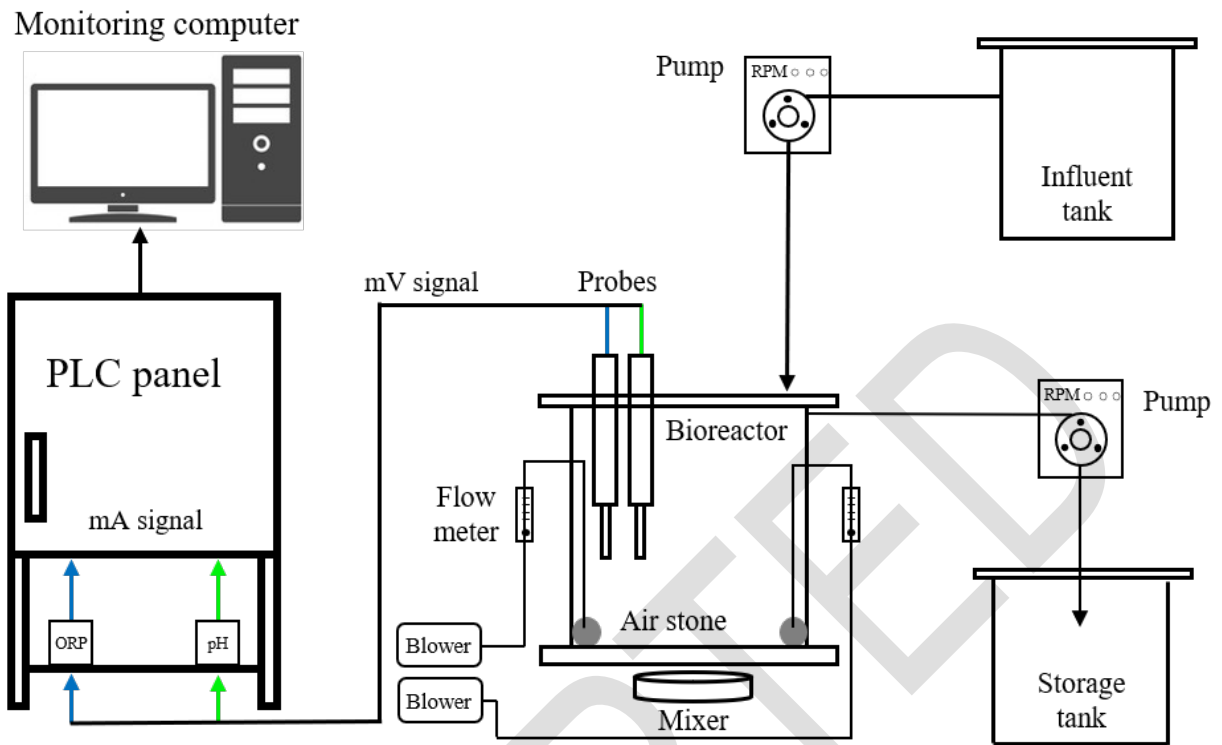


Fig. 1. Schematic design of liquid composting process and monitoring system

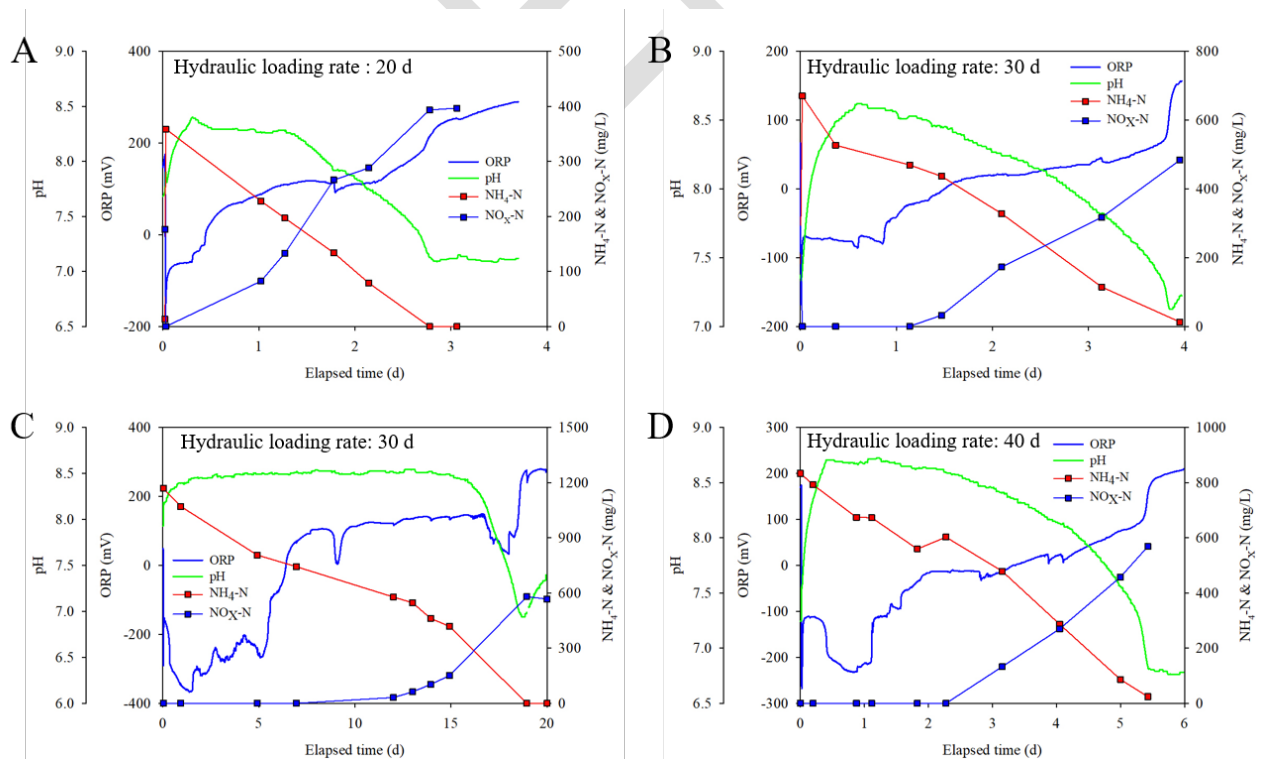
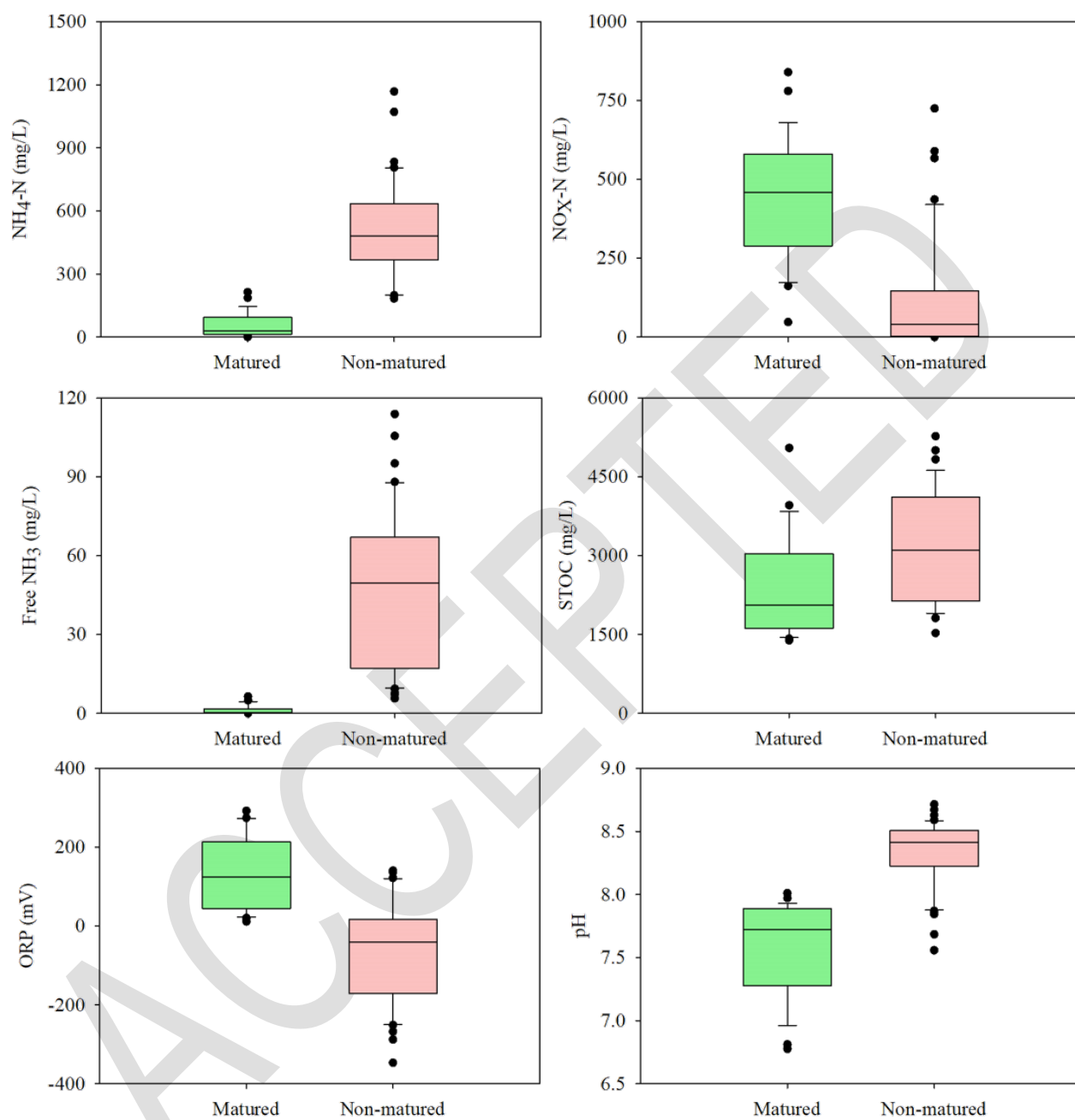


Fig. 2. Changes in ORP and pH profiles and soluble nitrogen concentrations during bioreactor operation

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Fig. 3. Comparison of chemical characteristics between matured and non-matured liquid compost

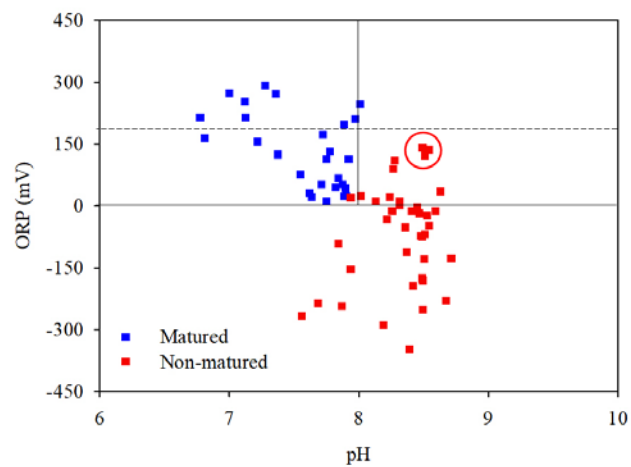


Fig. 4. Distribution of oxidation-reduction potential (ORP) and pH in matured and non-matured liquid compost.

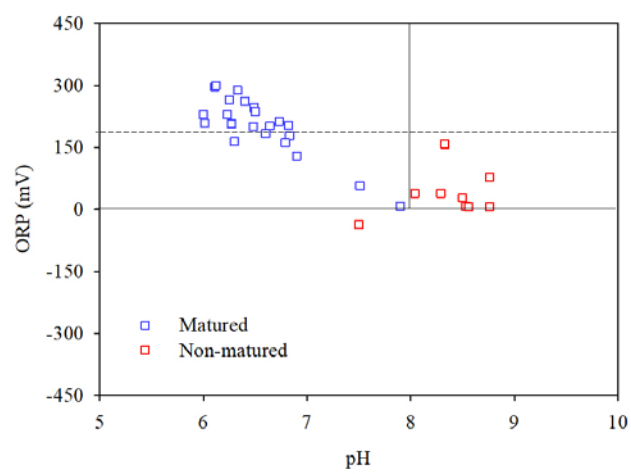


Fig. 5. Field validation of liquid compost maturity classification using oxidation-reduction potential (ORP) and pH in combination